

Responses of microbial community to soil warming in warm-temperate evergreen broad-leaved forests

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To elucidate the effect of climate warming on the soil heterotrophic microbial community in warm-temperate, evergreen broad-leaved forests, we conducted a soil-warming experiment in a secondary forest located in the city of Higashi-Hiroshima in western Japan. We established ten trench plots (1 m × 1 m) with root barriers to prevent root regrowth in the forest. The plots were divided into warming and control treatments. Infrared heaters were used to increase the soil temperature of the warming plots by about 2.5°C for three years. We used phospholipid fatty acid (PLFA) analysis to examine the composition of the soil heterotrophic microbial community. There were no significant differences in the total content of PLFAs (TotPLFAs) and fungal PLFAs (FungPLFAs) between the warming and control plots. However, warming caused an increase in the amount of bacterial PLFAs (BactPLFAs), the result being a lower ratio of FungPLFAs to BactPLFAs (F/B ratio) in the warming plots. In addition, PLFAs characteristic of gram-negative bacteria increased in the warming plots. The results indicated that the soil heterotrophic microbial community in this warm-temperate, evergreen broad-leaved forest was sensitive to climate warming.

Key Words : evergreen broad-leaved forest, microbial community composition, phospholipid fatty acid (PLFA), soil warming, warm-temperate zone

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1. Introduction

The global mean temperature increased by about 0.6°C during the 20th century as a result of climate warming and is expected to increase by another 1.1–6.4°C by the end of the 21st century (IPCC, 2007). Terrestrial ecosystems contain almost three times the amount of carbon found in the atmosphere, with forests comprising by far the largest fraction (Schlesinger, 1997; IPCC, 2001). Because soils contain more than two-thirds of the carbon in forest ecosystems (IPCC, 2001), understanding the effect of climate warming on the functionality of terrestrial ecosystems requires knowledge about how forest soil processes respond to soil warming. Soil warming-induced changes in the composition and biomass of soil microbial communities have the potential to cause changes in soil carbon flow and are therefore likely to affect the global carbon cycle (Schimel and Gullledge, 1998; Bardgett *et al.*, 2008).

To date there have been several experimental studies of the effects of soil warming on the biomass and composition of the soil microbial community in different types of forests

(Schindlbacher *et al.*, 2011; Frey *et al.*, 2008; Allison and Treseder, 2008; Kuffner *et al.*, 2012). The results of these studies have varied widely: no change in microbial biomass and composition in a spruce forest (Schindlbacher *et al.*, 2011); reduction in fungal biomass and shift in the microbial community in a mixed deciduous forest (Frey *et al.*, 2008); and increases in bacterial and fungal biomass in a conifer forest (Allison and Treseder, 2008). These results suggest that the effects of future climate warming on the biomass and composition of the soil microbial community will differ among forest ecosystems with various tree species, locations and climates.

Evergreen broad-leaved forests are the dominant natural forest type in the warm-temperate zone of the east coasts and adjacent islands of Asia (Kira, 1978). This type of forest has a much higher carbon cycling rate than forests in cool climates, such as cool-temperate and boreal forests (Suzuki, 1982). Wang *et al.* (2012) recently used open-top chambers to examine the effects of climate change on soil respiration in experimental stands of *Quercus glauca* and reported that elevated temperature treatments caused a significant change

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in the temperature sensitivity (Q_{10}) of heterotrophic respiration, a result that might have been caused by a change in the structure of the microbial community. However, there is little available information relevant to the potential impact of climate warming on soil microbial communities in warm-temperate, broad-leaved forests.

In this study we conducted a field experiment to examine the effect of soil warming on the soil heterotrophic microbial community in a warm-temperate, evergreen broad-leaved forest. We used infrared heaters to increase the soil temperature of experimental plots in a forest stand for three years and used phospholipid fatty acid (PLFA) analysis to examine changes in the heterotrophic microbial community.

2. Materials and methods

The experiment was conducted in a 25-year-old secondary evergreen oak forest following the pine tree death event occurred in 1980s in the central part of the city of Higashi-Hiroshima in western Japan ($34^{\circ}41'N$, $132^{\circ}72'E$; 335 m a.s.l.). The forest area is approximately 1.6 km². The forest is dominated by *Quercus glauca*, interspersed with *Symplocos kuroki*, and *Ilex pedunculosa*. In 2007, the tree density and canopy height being 1,450 stems ha⁻¹ and about 10 m, respectively. The soils are classified as coarse-textured, residual, immature soils with a parent rock of granite (Japan National Land Survey Division, Land and Water Bureau).

Ten plots on the forest floor within a circular area with a diameter of 40 m were randomly chosen. In order to prevent root regrowth into the plots, all the plots were treated by the trenching method developed by Liang *et al.* (2010). Ten 1 × 1 m root exclusion plots were established. Trenches with a width of 0.5 to 1 cm were dug down to 50 cm along the plot boundaries using a root-cutting chainsaw (CSVN671AG, Kioritz Co. Ltd., Tokyo, Japan) and then polyvinyl chloride (PVC) sheets (4 mm thick) were installed in the trenches to a depth of 50 cm to prevent root penetration. The plots were randomly allocated to two groups, five of which, the warming plots, were warmed continuously for 24 hours with infrared heaters beginning in October 2007. The other five served as controls. In each warming plot a 58-cm-long, 800-W infrared heating lamps were suspended in the middle of the plot at 1.6 m above the ground (Aguilos *et al.*, 2013). The soil temperature and moisture content of each plot were measured continuously at soil depths of 5 and 15 cm. Data were stored on data-loggers at 10 s intervals (Liang *et al.*, 2010).

Soil samples for PLFA analysis were collected from the mineral soil (0–5 cm) of the warming and control plots in October 2010. Three cores (5 cm diameter × 5 cm deep) were taken from each plot. One core was taken from the central point, and the other two cores were taken from the 1/4 point and 3/4 point of the diagonal line in each 1 × 1 m plot. We mixed the three cores from each plot to obtain one composite sample. The samples were returned to the laboratory, freeze-dried, sieved (< 2 mm) to remove gravel and dead roots, and stored at $-80^{\circ}C$ until PLFA analysis.

Lipids were extracted by using a Bligh and Dyer (1959)

extraction, as modified by White *et al.* (1979) and Frostegård *et al.* (1993). Briefly, an aliquot of 1–2 g (dry weight) of soil was extracted with a chloroform-methanol-citrate buffer mixture (1:2:0.8). The lipids were separated into neutral lipids, glycolipids, and phospholipids on a silicic acid column (Sep-PakTM plus silica; Waters Corp., Milford, MA, USA) (Arao *et al.*, 2001). The phospholipids were esterified with a HCl-Methanol Reagent (Tokyo Kasei Kogyo Co. Ltd., Tokyo, Japan) (Stoffel *et al.*, 1959). The resulting fatty acid methyl esters were separated with a gas chromatograph (GC-2014; Shimadzu Corp., Kyoto, Japan) equipped with a capillary column (30 m DB-5 ms, Phenyl-Methyl/Silicone; J&W Scientific Inc., Folsom, CA, USA). Helium was used as a carrier gas. Peak areas were quantified by adding methyl nonadecanoate fatty acid (19:0) as an internal standard. The fatty acid nomenclature described by Frostegård *et al.* (1993) was used. The total content of PLFAs (TotPLFAs) was used as a metric of total microbial biomass (Frostegård *et al.*, 1993). The fatty acids i15:0, a15:0, 15:0, i16:0, 17:0, i17:0, cy17:0, 18:1 ω 7c, and cy19:0 were chosen to represent bacterial PLFAs (BactPLFAs) (Frostegård *et al.*, 1993; Federle, 1986). The amount of 18:2 ω 6,9 was used as a metric of fungal biomass (FungPLFA) (Federle, 1986; Frostegård and Bååth, 1996; Xue *et al.*, 2008).

PLFAs are classifiable into several categories according to their molecular structure (Yoshitake *et al.*, 2006). In this study PLFAs were classified into five categories (straight chain saturated fatty acids, branched chain fatty acids, unsaturated fatty acids, cyclopropyl fatty acids (cy17:0, cy19:0 and 18:2 ω 6,9) and summed their mole percents. Cyclopropyl fatty acids and branched chain fatty acids are known to be characteristic of gram-negative bacteria and gram-positive bacteria, respectively (Federle, 1986). Previous studies have shown that this classification is useful for detection of shifts in microbial community structure (Yoshitake *et al.*, 2006).

Student's *t*-test was used to test for significant differences in soil properties and PLFA content between the warming and control plots. All statistical analyses were carried out with SPSS software (v. 19.0; SPSS Inc., Chicago, IL). We considered differences to be significant if the type I error rate (*P*) was 5% or less.

3. Results and discussion

Warming consistently increased soil temperatures at 5 cm depth by about 2.5°C throughout the three-year experimental period. The ranges of daily average soil temperature were 4.1–22.6°C and 7.1–25.8°C in the control and warming plots, respectively. The corresponding ranges of soil moisture content were 20.5–37.6% and 20–36.5%, respectively. The soil moisture content of the warming and control plots did not differ significantly (Student's *t*-test, *P* > 0.05) during the three-year experimental period.

The TotPLFA content was not significantly affected by the three years of experimental soil warming (Student's *t*-test, *P* > 0.05, Table 1). A similar insensitivity of microbial

Table 1. Fatty acid content and the proportion of fatty acids in the microbial communities from the warming and control plots after three years. Mean and standard deviation (in parentheses, n = 5) are given.

	Warming	Control	<i>P</i>
Fatty acid content (nmol g ⁻¹)			
TotalPLFA	376.2 (86.6)	378.7 (84.6)	0.965
BactPLFA	115.4 (17.7)	90.0 (7.0)	0.017*
FungalPLFA	21.6 (8.9)	25.7 (7.9)	0.467
F/B ratio	0.2 (0.06)	0.3 (0.07)	0.040*
Fatty acid Proportions (mole percent)			
Straight chain saturated fatty acids	25.3% (5%)	23.9% (2%)	0.560
Branched chain fatty acids	28.4% (6%)	26.1% (2%)	0.442
Cyclopropyl fatty acids	12% (6%)	5.5% (2%)	0.047*
Unsaturated fatty acids	37.7% (3%)	40% (6%)	0.439
18:2ω6,9	5.7% (2%)	6.8% (1%)	0.309

Significant differences between the warming and control estimated by Student's *t*-test.

*, *P* < 0.05

biomass to soil warming has been reported in several previous studies (Schindlbacher *et al.*, 2011; Zhang *et al.*, 2005; Zogg *et al.*, 1997), although Petersen and Klug (1994) reported some change in soil microbial biomass in response to soil warming.

Bacterial and fungal biomass responded differently to soil warming: a significant increase in BactPLFA content occurred in the warming plots (Student's *t*-test, *P* < 0.05, Table 1), whereas FungPLFA content was not significantly affected by the three years of soil warming. As a result, the ratio of fungal to bacterial biomass decreased significantly (Table 1), an indication of a significant shift in the composition of the microbial community.

Differences in the relative abundance of cyclopropyl fatty acids, which are characteristic of gram-negative bacteria (Harwood and Russell, 1984), provided further evidence of a temperature-induced shift in community composition (Student's *t*-test, *P* < 0.05, Table 1). In addition, some new fatty acids (15:0, 17:0, 17:1ω7, 18:0, 20:4ω6, and 20:5ω3) were detected in the warming plots, whereas the fatty acid 16:1 became undetectable (Fig. 1), a further indication of changes in the composition of the microbial community.

Results of previous soil-warming field studies of microbial communities in forest ecosystems have been rather inconsistent though bacterial biomass was generally unaffected by elevated temperatures (Table 2). Schindlbacher *et al.* (2011) reported that a 4°C increase in soil temperature during the snow-free season had no influence on the microbial community composition in a temperate, mountain forest soil. In contrast, Frey *et al.* (2008) reported a significant re-

duction in total microbial biomass and a shift in microbial community composition in a mixed deciduous forest after 12 years of soil warming of 5°C above the ambient temperature. In a study of an Alaskan boreal forest, Allison and Treseder (2008) reported that bacterial and fungal abundance declined by more than 50% in a closed-top greenhouse treatment. However, because their warming treatment caused drying of the soil, the observed change in microbial biomass may have been partly due to the change in moisture content. Arnold *et al.* (1999) indicated that soil microbial response to experimental soil warming was most evident when soil moisture was not limiting (range between 20% and 120%). In this study the soil moisture content did not differ significantly between treatments, and the observed shift in the microbial community can therefore be attributed solely to the effect of warming. Considering the relatively short duration (3 years) of our study and the lower warming temperature (+2.5°C), we conclude that the soil bacterial community in this warm-temperate, evergreen broad-leaved forest was remarkably sensitive to warming effects.

It has been suggested that bacterial community composition influences heterotrophic respiration and can potentially influence soil carbon storage (Cleveland *et al.*, 2007). Although studies have yet to be carried out of the impact of climate change on members of the soil microbial community that are symbiotic with plants, these results suggest that changes in the heterotrophic microbial community are among the mechanisms through which climatic warming affects the carbon cycle of warm temperate evergreen broad-leaved forest.

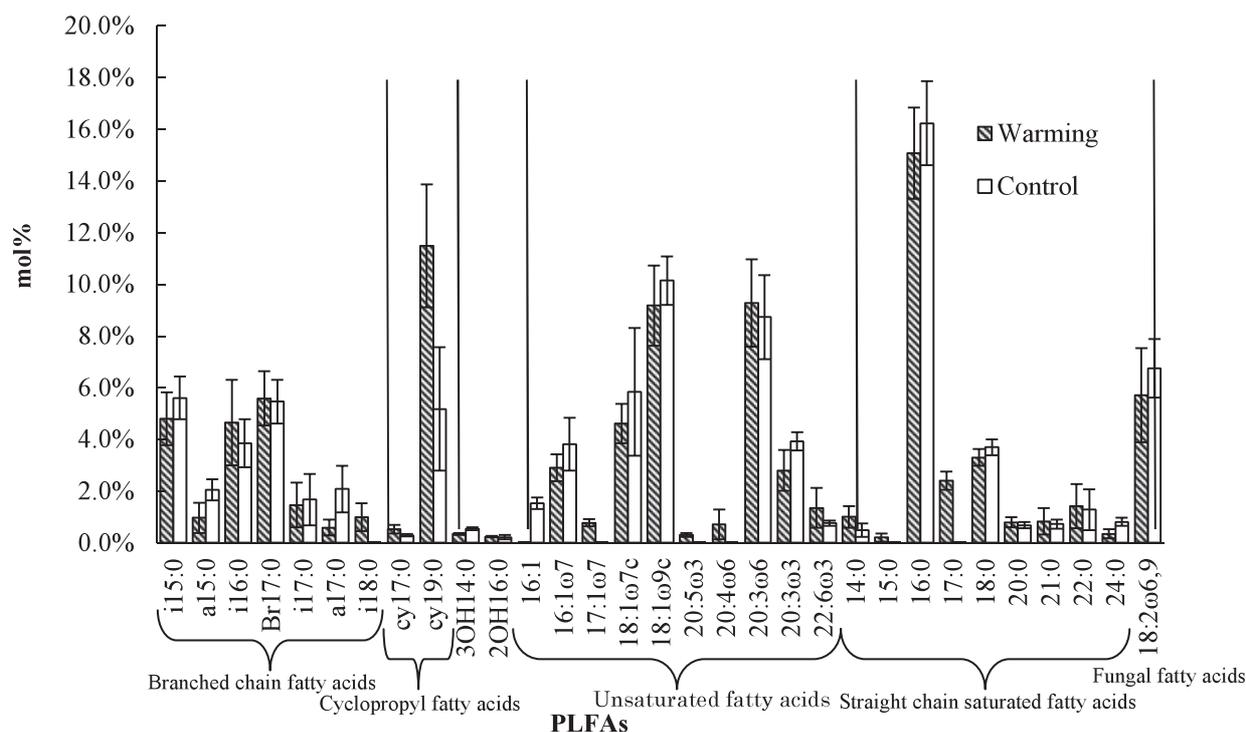


Fig. 1. Changes in the mole fraction of selected phospholipid fatty acids (PLFAs) extracted from the soils of the control and warming plots. Error bars are one standard error of the mean of five plots. W: warming; C: control.

Table 2. Effects of elevated temperature on soil microbial communities in experimental forest stands

Species	Temperature increase (°C)	Exposure period	Biomass			Reference
			Total	Bacteria	Fungi	
Conifer forest						
<i>Picea mariana</i>	+0.5 (Soil T)		-	-		Allison and Treseder (2008)
<i>Picea mariana</i> spruce-fir	+0.4-0.9	3 years	-	ns	ns	Bergner <i>et al.</i> (2004) Arnold <i>et al.</i> (1999)
Deciduous broad-leaved forest						
<i>Acer saccharum</i>	5,15,25	16 weeks	-			Zogg <i>et al.</i> (1997)
<i>Quercus velutina</i>	+5	12 years	-			Frey <i>et al.</i> (2008)
<i>Acer rubrum</i>	+5	12 years	-			Frey <i>et al.</i> (2008)
<i>Betula papyrifera</i>	+5	12 years	-			Frey <i>et al.</i> (2008)
<i>Acer pensylvanicum</i>	+5	12 years	-			Frey <i>et al.</i> (2008)
Mixed forest (deciduous broad-leaved trees and conifers)						
<i>Picea abies</i>	+4	6 years	ns	ns	ns	Schindlbacher <i>et al.</i> (2011)
<i>Fagus sylvatica</i>	+4	6 years	ns	ns	ns	Schindlbacher <i>et al.</i> (2011)
<i>Abies alba</i>	+4	6 years	ns	ns	ns	Schindlbacher <i>et al.</i> (2011)
<i>Picea abies</i>	+4	4 years		ns		Kuffner <i>et al.</i> (2012)
<i>Fagus sylvatica</i>	+4	4 years		ns		Kuffner <i>et al.</i> (2012)
<i>Abies alba</i>	+4	4 years		ns		Kuffner <i>et al.</i> (2012)
Evergreen broad-leaved forest						
<i>Quercus gluaca</i>	+2.5	3 years	ns	+	ns	Present study

+ increased; - decreased; ns not significant

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